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## An affordable methodology for quantifying waterborne microplastics - an emerging contaminant in inland-waters

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### ABSTRACT

The occurrence of microplastics in marine habitats is well documented and of growing concern. The presence of these small (<5 mm) pieces of plastic is less well recorded in inland water systems. In this paper, we determine a cost-efficient and straightforward method for the collection and identification of microplastics in UK inland waters. We found pieces of microplastic from all sample sites ranging from over 1000 L<sup>-1</sup> in the River Tame, to 2.4 L<sup>-1</sup> in Loch Lomond. The presence of microplastics in all waters tested suggests it should now be classed as an emergent contaminant, with routine monitoring required.

### INTRODUCTION

Microplastics are pieces of plastic less than 5 mm in size (Thompson *et al.*, 2009) that come from a variety of either primary or secondary sources. Primary sources are those plastics purposefully manufactured to such a size for use in cosmetics or cleaning products, or as part of the general plastic production system. Secondary sources of microplastics are those fragments of plastic produced through the breakdown of larger pieces. Both types can enter inland water systems through a variety of ways and the full impacts of these pellets, fragments and fibres on ecosystems, wildlife and indeed our own health are not yet fully understood (Cole *et al.*, 2011; Eerkes-Medrano *et al.*, 2015; de Souza Machado, 2017).

Some pollutants in inland waters are regularly monitored and guidelines are enforced to ensure levels do not exceed beyond stated “safe” concentrations. Other pollutants – so called emergent contaminants - such as

pharmaceutical waste, personal care products and illicit drugs are only just being recognised as issues, and work is being conducted to investigate these problems fully (Tran *et al.*, 2018). In this paper, we use a practical and inexpensive method to highlight the widespread presence of microplastics in UK mainland waters, indicating it is essential they are now considered an important emergent contaminant. The methodology described purposefully only uses standard laboratory equipment and a commercially available fluorescence light attachment, offering the potential of this method being used to acquire data on many more sites on a regular basis, by a wide variety of organisations and collaborators.

### METHODS

Four, clean, one-litre glass amber bottles with plastic lids, with a standard (~ 2.5 cm) opening, were rinsed thoroughly with water from the sample site (see Tab. 1 for a full list of the sites). Each bottle was then filled to the very top with site water from a depth of approximately 5 - 10 cm from the surface, and capped underwater (Green *et al.*, 2018). The sampler remained downstream of the bottle being filled at all times, and water was collected from a safe wading distance and sampling was conducted between June 2018 and February 2019. Samples were stored in the laboratory at 4°C in the dark, until analysis was completed.

The contents of each bottle were filtered using a glass vacuum-pump filtering system through a GF/C glass filter (1.2 µm pore size; GE Healthcare Whatman™) - chosen due to their relative affordability. Filters were dried and analysed for microplastic numbers and types (Hidalgo-Ruz *et al.*, 2012; Ravit *et al.*, 2017) using a dissecting microscope (with a magnification level ranging from 10 to 40) using either a standard visible light system or a

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fluorescence lighting system attached to the same microscope. For this research a NIGHTSEA™ Stereo Microscope Fluorescence Adapter was used. The designation was Royal Blue (RB) with an excitation of 440-460 nm and emission filter of 500 nm (longpass). This set-up was chosen as the adaptor fits to any standard dissection microscope and is low-cost (~1000 USD) compared to specialised fluorescence microscope set-ups used in previous studies (Qui *et al.*, 2015). Comparisons were conducted to determine which lighting system (visible or fluorescence) made it easier to detect microplastics. When particles could not be visually identified as microplastic suspected pieces were tested using the bending test (to see if they snapped or, if they were plastic, bent) and hot-needle technique (Hurley *et al.*, 2018). The latter involves placing a very hot needle or pin near a suspected piece of plastic. If it is plastic, it will melt or curl.

Measures were taken to minimise the contamination of water and filters throughout the procedure. Filters were kept in glass petri-dishes and only uncovered for the necessary analysis, such as the hot-pin method. Lab coats were regularly cleaned with lint removers. A series of control samples were made by using the same type and manufacturer of bottles and lids, following the usual filtering process but using pre-filtered water – where necessary, results from these controls were subtracted from the experimental samples to give a set of normalised results. These controls were conducted during any filtering and sampling period in the laboratory. Additional precautions to ensure robustness of the controls were also adhered to, such as leaving the vacuum pump system on for approximately the same time as it took the inland water samples to go through the filter.

Each sample was counted four times by a trained observer, and the samples were counted in a randomised order to help prevent bias. An average of each count was taken as the result for that sample. For samples that contained many pieces of microplastic, when observers had counted 1000 single pieces of microplastic (all types combined) they stopped counting and the result was recorded as “>1000”.

Due to the equipment used and identification process implemented all microplastic pieces counted were longer than approximately 200  $\mu\text{m}$  and wider than around 50  $\mu\text{m}$ , or had an area greater than approximately 2500  $\mu\text{m}^2$ .

Significant differences between results were determined by an independent samples *t*-test, using IBM SPSS Statistics 25, after checking the usual assumptions of parametric tests were not violated.

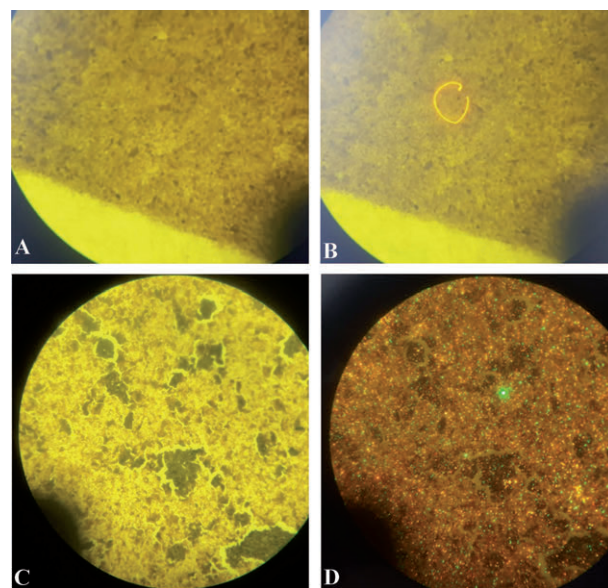
Using the fluorescence lighting system attached to the dissecting microscope produced greater detection levels of microplastics than just using a standard, visible light system; especially for identifying fragments. For instance,

there was a statistically significant difference when looking at fragments under the two different lighting conditions for Afon Cegin (the site used for most of the method development):  $t(6.27) = 5.07$ ,  $P=0.001$ . The average count for microplastics in one litre of water observed under the standard lighting system was 22.3 (standard error,  $SE = 3.4$ ) compared to a mean average of 73.7 ( $SE = 9.5$ ) using fluorescence lighting.

The advantage of using fluorescence can be seen in Fig. 1: the microplastic fragment cannot be observed on the filter paper in Fig. 1A, taken under normal lighting conditions, whilst it is clearly visible on the same filter paper in Fig. 1B for which the fluorescence system is used (the same for Fig. 1 C,D). However, as some organics and minerals present in the samples also fluoresce, it is vital that switching between the two lighting sources is considered and standard techniques, *e.g.* bending test, hot-pin method, Raman spectroscopy (Araujo *et al.*, 2018) are used to be sure of microplastic identification.

## RESULTS

All sites analysed had microplastics present (Tab. 1). Loch Lomond had the lowest number of microplastics with a total of 2.4 pieces per litre ( $\text{L}^{-1}$ ) while River Tame had the most with over 1000 pieces  $\text{L}^{-1}$ . As with all reported concentrations of microplastics this is the normalised result, following the subtraction of a set of pre-filtered control, or ‘blank’, samples. However, to



**Fig. 1.** A section of filter paper looked at through x40 magnification for microplastic identification. a,c) Photographed using only standard, visible light; b,d) photographed using a fluorescence lighting system.

**Tab. 1.** Numbers and types of microplastic found in inland waters from the UK. Results are a mean average ( $n = 4$  for all sites except Ullswater and Afon Cegin,  $n = 6$ ) from one litre of site water. Microplastics were categorised as fibre (pieces of line or filament), fragment (pieces broken off from larger plastics), film (breakdown from bags, wrappers etc), pellet (microbeads and nurdles), or foam (broken pieces of polystyrene items).  $\pm$  indicates standard error (there is no standard error for River Tame due to the count being recorded as  $>1000$  – other microplastic types may also have been present but the methodology dictates counting stops after 1000 individual microplastic pieces are detected). Numbers in parentheses indicate mean averages ( $n = 4$ ) of procedural blanks, consisting of pre-filtered water. These averages were subtracted from the all relevant sample counts before averaging and statistical analysis.

Site	Location	Microplastic type					Total
		Fibre	Fragment	Film	Pellet	Foam	
River Thames	51°30'30.7"N 0°06'37.0"W	6.9 $\pm$ 1.5(0.25)	74.4 $\pm$ 11(1.75)	1.1 $\pm$ 0.4	0.1 $\pm$ 0.1	1.7 $\pm$ 0.9	84.1
Chester reedbed	53°12'28.6"N 2°54'12.0"W	1.8 $\pm$ 1.9	4.3 $\pm$ 0.3(0.25)	0.1 $\pm$ 0.1	0.4 $\pm$ 0.4	1.1 $\pm$ 0.2	7.6
Ullswater	54°34'30.4"N 2°54'29.4"W	5 $\pm$ 0.5(0.75)	14 $\pm$ 1.4(1)	3.3 $\pm$ 0.4	4.9 $\pm$ 0.9	2.5 $\pm$ 0.4	29.5
River Irwell	53°29'19.2"N 2°16'07.9"W	0(1.6)	84.8 $\pm$ 31.7(13.1)	0(0.1)	0(0.1)	0	84.8
River Tame	53°27'44.6"N 2°06'03.9"W	0(1.6)	>1000(13.1)	0(0.1)	0(0.1)	0	>1000
River Blackwater	51°43'34.9"N 0°45'23.7"E	3 $\pm$ 0.2	10.7 $\pm$ 3(1.5)	0	0	1.4 $\pm$ 0.8	15.1
Falls of Dochart	56°27'45.2"N 4°19'13.2"W	1.1 $\pm$ 0.5	2.2 $\pm$ 0.7(0.4)	0	0	0	3.3
Loch Lomond	56°06'43.9"N 4°37'25.8"W	0.9 $\pm$ 0.4	1.5 $\pm$ 0.6(0.3)	0	0	0	2.4
Afon Cegin	53°13'53.3"N 4°06'39.4"W	14.8 $\pm$ 5.7(16)	49.7 $\pm$ 9.5(24)	5.7 $\pm$ 3.3(1)	2.7 $\pm$ 1.1(1)	4 $\pm$ 3.4(5)	76.9
Llyn Cefni	53°16'12.4"N 4°20'22.4"W	7.4 $\pm$ 1.1(0.5)	16.8 $\pm$ 4.1(1)	7.7 $\pm$ 1.2	8.5 $\pm$ 2	2.9 $\pm$ 1.4	43.2

ensure validity of the method it was first checked to confirm the non-normalised results were statistically different to the control samples *e.g.* for Loch Lomond there was a mean average of 2.4 (SE = 0.3) pieces of microplastic fragment in the raw samples, and an average of 0.3 (SE = 0.5) in the controls; showing a statistically significant difference  $t(21.45) = 3.99$ ,  $P < 0.001$ .

## DISCUSSION

With every sample tested showing evidence of plastics – the most common type across all sites being fragments – it is suggested microplastic pollution is now endemic across all inland water systems in mainland UK. We found it in major rivers running through large urban regions, such as the Thames, Tame and Irwell, as well as remote rivers (Falls of Dochart); wetlands (Chester reedbed); lakes and lochs (Ullswater and Lomond), and reservoirs (Cefni). Some of these are iconic British water systems, and they now all contain microplastic pollution.

There are other methods for measuring microplastic concentrations from the environment and these can have advantages. However, our results illustrate a low-cost, low-tech method for sampling and quantifying microplastic contamination. The process is also relatively time efficient, with it taking approximately 10 to 30 minutes to count the plastic particles visible on most filter papers. There are clear limitations to our affordable and efficient methodology, and not all types of plastic will fluoresce under such conditions (Qiu *et al.*, 2015), so eventually standardised spectroscopic and/or

chromatographic methods (imaging FT-IR, microscopy, pyrolysis GC-MS) may become available, and affordable, for higher throughput of microplastic analysis in environmental samples. However, until that time regular monitoring of water systems using the methods outlined in this paper should become routine, as our findings suggest microplastics are now an emergent contaminant.

## CONCLUSIONS

Further work is now essential to investigate fully the health risks of microplastics – to humans and ecosystems – so that “safe” levels can be ascertained, and removal or mitigation processes can be put in place. This could involve the development and use of ecological engineering initiatives such as specially designed constructed treatment wetlands (CTWs) to filter-out plastic particles.

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